# Effective number of breeding adults in Oregon spotted frogs (Rana pretiosa): genetic estimates at two life stages 

Ivan C. Phillipsen • Jay Bowerman • Michael Blouin

Received: 29 October 2008 / Accepted: 10 February 2009
© Springer Science+Business Media B.V. 2009


#### Abstract

We used genetic methods to estimate the effective number of breeders $\left(N_{\mathrm{b}}\right)$ in a population of Rana pretiosa, an imperiled amphibian in western North America. Microsatellite data was gathered from large samples of adults, eggs, and juveniles collected in 2006. We wished to determine where in the life cycle the greatest reductions in $N_{\mathrm{b}}$ occur, and to compare genetic estimates of $N_{\mathrm{b}}$ to an egg mass count estimate of the number of breeding adults. We predicted that $N_{\mathrm{b}}$ estimated at the metamorph stage would be reduced by increased variance in family size due to egg mass mortality. Contrary to our prediction, estimates of $N_{\mathrm{b}}$ at the egg and metamorph stages were similar. Thus, we found no evidence of inflated variance in family size between the two stages. If our results for this population are typical for $R$. pretiosa, then increased variance in family size during the egg to metamorph stage may not be a strong factor in reducing the effective population sizes $\left(N_{\mathrm{e}}\right)$ relative to the census sizes $(N)$ in this species.


Keywords Amphibians • Anura • $N_{\mathrm{e}} / N$ ratio •
Temporal method $\cdot$ Linkage disequilibrium $\cdot$
Microsatellites

[^0]
## Introduction

Effective population size $\left(N_{\mathrm{e}}\right)$ is a fundamental parameter in the theory and practice of conservation genetics. Related to $N_{\mathrm{e}}$ is the effective number of breeders, $N_{\mathrm{b}}$, a parameter influenced by most of the same demographic factors as $N_{\mathrm{e}}$ but which applies to only the breeding adults of a population in a single reproductive season. Estimates of $N_{\mathrm{b}}$ or $N_{\mathrm{e}}$ in natural populations are usually much lower than the census population size, $N$ (e.g., Frankham 1995). What causes $N_{\mathrm{e}}$ and $N_{\mathrm{b}}$ to be lower than $N$ is not well understood for many species.

The ongoing loss of global amphibian diversity is a widely recognized ecological crisis (Stuart et al. 2004). Values of $N_{\mathrm{e}} / N$ and $N_{\mathrm{b}} / N$ reported for amphibians range widely, from 0.001 (Easteal 1985) to $>0.7$ (Brede and Beebee 2006). What features of the life histories of different species might predispose them to have different ratios? For example, there is some intriguing evidence that toads of the genus Bufo have $N_{\mathrm{e}} / N$ ratios an order of magnitude lower than those of frogs of the genus Rana (Hoffman et al. 2004; Brede and Beebee 2006). Understanding what factors in the life cycle of amphibians are most responsible for reductions in $N_{\mathrm{b}}$ or $N_{\mathrm{e}}$ could be very useful for managing loss of genetic diversity in these taxa.

The two factors thought to most dramatically reduce $N_{\mathrm{e}}$ in animal populations are fluctuating population size and non-random variance in family size (Frankham 1995). Pond-breeding frogs may be particularly susceptible to reductions in $N_{\mathrm{e}}$ by these factors. Populations of frogs in the family Ranidae often go through "boom and bust" cycles from year to year as a result of the environmental instability of their breeding habitats (Berven 1995). In addition, variance in family size for these frogs may be greater than under random (i.e., Poisson distributed)
expectations due to the loss or survival of whole families during the egg stage of the life cycle (Crow and Morton 1955; Rowe and Beebee 2004). Entire egg masses or portions of egg masses are often lost to desiccation, freezing, predation, or disease (Briggs 1987; McAllister and Leonard 1997). If survival operates at the family level, then the inflation of variance in family size (and reduction in $N_{\mathrm{b}}$ ) can be enormous (Crow and Morton 1955). In this study we focus primarily on reduction in $N_{\mathrm{b}}$ incurred during the egg to metamorph stage.

The number of breeding individuals in a given year is often estimated for ranid frog populations by doubling the number of discrete egg masses found in the pond(s) that year (Crouch and Paton 2000). This estimate is sometimes used to estimate $N_{\mathrm{b}}$ (Merrell 1968; Berven and Grudzien 1990; Watson et al. 2000). Estimating $N_{\mathrm{b}}$ this way assumes that each female lays only one egg mass per year, each egg mass is fertilized by a single male, each male breeds with only one female per year, and that family size is poisson distributed. The first three assumptions are likely to hold for 'explosive breeding' species, which engage in a single, brief (e.g., 1-3 nights) reproductive bout each year (Wells 1977). The fourth assumption is much more dubious, but how much reduction in $N_{\mathrm{b}}$ results from non-random survival between egg laying and metamorphosis has not been estimated.

Here we used genetic estimates of $N_{\mathrm{b}}$ in a population of the Oregon spotted frog (Rana pretiosa) to estimate the reduction in $N_{\mathrm{b}}$ owing to reproductive strategy and to nonrandom survival among families. We analyzed molecular genetic data from large samples of adults, eggs, and postmetamorphic juveniles collected during a single season (Fig. 1). We estimated $N_{\mathrm{b}}$ at two stages in the life cycle using variances in microsatellite allele frequencies between: (1) adults and eggs; and (2) adults and metamorphs. This is a single-season version of Waples' (1989) temporal method of $N_{\mathrm{e}}$ estimation, and our approach is similar to that of Scribner et al. (1997).

Given our field observations and the fact that R. pretiosa is an explosive breeding species, our a priori expectation was that neither extra-pair fertilization nor multiple mating has a strong influence on $N_{\mathrm{b}}$ in this species. Thus, the estimate of the effective number of breeders $\left(\hat{N}_{\mathrm{b}}\right)$ derived from the allele frequency differences between adults and eggs should be similar to the egg mass count estimate of the actual number of breeders $\left(\hat{N}_{\mathrm{ab}}=2 \times\right.$ number of egg masses $)$ On the other hand, mortality of all or parts of some egg masses is well documented in our and other populations of $R$. pretiosa (Bowerman, personal observation; Licht 1971). Non-random survival among individuals due to egg mass mortality (i.e., family-correlated survival) would reduce $\hat{N}_{\mathrm{b}}$ as measured by allele frequency differences between adults and metamorphs. Therefore, our prediction was that the adultmetamorph $\hat{N}_{\mathrm{b}}$ would be much less than the adult-egg $\hat{N}_{\mathrm{b}}$.

$\hat{N}_{\mathrm{ab}} \approx \hat{N}_{\mathrm{b}}{ }^{\text {TME99 }}$ if there are no multiple matings or extra pair fertilizations
$\hat{N}_{\mathrm{ab}}>\hat{N}_{\mathrm{b}}{ }^{\text {TMMea }}$ if egg mass mortality results in family-level survival

$$
\text { Prediction: } \hat{N}_{\mathrm{ab}} \approx \hat{N}_{\mathrm{b}}^{\text {TM-Egg }}>\hat{N}_{\mathrm{b}}^{\text {TM-Meta }}
$$

Fig. 1 Sampling scheme for estimating the effective number of breeders $\left(N_{\mathrm{b}}\right)$ in a population of $R$. pretiosa. Three samples were collected in 2006: adults, eggs, and metamorphs. For each sample, allele frequencies were calculated for seven microsatellite loci. Two estimates of $N_{\mathrm{b}}$ were derived from this genetic data using the temporal method. The first was based on allele-frequency differences between the adult and egg samples $\left(\hat{N}_{\mathrm{b}}^{\mathrm{TM}-E g g}\right)$, while the second was based on differences between the adult and metamorph samples ( $\left.\hat{N}_{\mathrm{b}}^{\mathrm{TM}}{ }^{\text {-Meta }}\right)$. An estimate of the actual number of breeding adults was calculated as twice the number of egg masses counted in the pond in $2006\left(\hat{N}_{\mathrm{ab}}\right)$. Our predictions as described in the text are represented by the relationships among $\hat{N}_{\mathrm{b}}^{\text {TM }}$ Egg,$\hat{N}_{\mathrm{b}}^{\text {TM }- \text { Meta }}$, and $\hat{N}_{\mathrm{ab}}$. Note that the 'TM' superscript used here refers to Waples' (1989) temporal moment method, but this sampling scheme and data were also used for the temporal likelihood (Berthier et al. 2002) analysis. Estimates of $N_{\mathrm{b}}$ were obtained separately for the egg and metamorph samples using the LD method (Hill 1981; Waples and Do 2007)

We also estimated $\hat{N}_{\mathrm{b}}$ from eggs and from metamorphs by the linkage disequilibrium (LD) method (Hill 1981). These estimates should be independent of the temporal method estimates (Waples 1991). Again, we predicted that the $N_{\mathrm{b}}$ estimate from the egg sample would be close to twice the number of egg masses $\left(\hat{N}_{\mathrm{ab}}\right)$, and that the estimate from the metamorph sample would be substantially less than the estimate from the egg sample.

Finally, we estimated $N_{\mathrm{e}}$ (as opposed to $N_{\mathrm{b}}$ ) in the adult population via the LD method and compared it to an estimate of $N$ obtained by intensive mark-recapture sampling. These data provide an additional point estimate of $N_{\mathrm{e}} / N$ for ranid frogs.

## Materials and methods

## Study organism

Oregon spotted frogs ( $R$. pretiosa) live in lakes and ponds in the Pacific Northwest, from southern Oregon in the United States to southern British Columbia in Canada (Hayes 1997; Nussbaum et al. 1983). R. pretiosa overwinter in permanent ponds or springs and breeding occurs soon after ice melt in the spring (Licht 1969; Leonard et al. 1997). During the $2-$ 4 week breeding season only mature adults are active at the surface, and the sex ratio is male biased (Watson et al. 2000; personal observations). Breeding is explosive, with most of the egg masses being deposited on one or a few nights (Licht 1969; McAllister and Leonard 1997). Females lay their eggs in communal piles in shallow water, and there may be several of these communal sites per pond. Boundaries between egg masses in a pile are very discrete for a week after laying, which makes counting and sampling individual masses straightforward.

The Oregon spotted frog has been extirpated from 70 to $90 \%$ of its original range (Hayes 1997). Fewer than 35 populations remain, and these are mostly small ( $N<1000$ ), isolated, and restricted to higher elevations (Hayes 1997; Cushman and Pearl 2007). R. pretiosa is a candidate for federal listing as endangered by the US Fish and Wildlife Service (2005), is considered "sensitivecritical" by the Oregon Department of Fish and Wildlife, and "endangered" by the state of Washington. It is an endangered species in Canada (Seburn and Seburn 2000). Thus, data on what controls $N_{\mathrm{e}}$ or $N_{\mathrm{b}}$ in this species could be useful for management of the remaining populations.

## Sample collections

Sampling took place in a pond located near Sunriver, Oregon ( $43.85018^{\circ} \mathrm{N}, 121.44768^{\circ} \mathrm{W}$ ). Adult and postmetamorphic juvenile frogs and were captured using underwater funnel traps (Gee's minnow traps) and dip nets. Adult frogs were individually marked with PIT tags. Metamorphs were not individually marked. Capturerecapture data from marked frogs was collected on 77 occasions from March 6th through December 9th 2006. For genetic sampling, a single toe clip was collected from each adult frog ( $n=208$ ) and from each metamorph sampled from the 2006 cohort $(n=401)$. Toe-clips were stored in Drierite desiccant (W. A. Hammond Drierite Co., Xenia, OH ). During the breeding season (late March through early April), the pond was carefully monitored for the presence of egg masses. About 45 egg masses were deposited on April 6th and were sampled within 48 h . We observed no egg mass mortality prior to taking our egg samples. Approximately ten eggs were sampled from each mass
( $n=452$ ). The eggs were allowed to develop for several days in the laboratory and then preserved in $70 \%$ ethanol. To our knowledge, no additional egg masses were deposited in 2006 and thus our sample of eggs included all families for that year. We excluded from our datasets any individual with missing data for one or more microsatellite loci. This resulted in an adult sample of 176, a total egg sample of 415 , and a metamorph sample of 308 .

The methods of $N_{\mathrm{b}}$ estimation used in this study assume samples are drawn at random (Hill 1981; Waples 1989). By collecting roughly ten eggs from each egg mass, we may have forced allele frequencies estimated from the egg sample to be more similar to the adult frequencies than if the same number of eggs had been sampled randomly from the entire pool of eggs produced in the pond (Waples, personal communication). This could result in an upward bias of the $N_{\mathrm{b}}$ estimates obtained using the egg sample. To avoid this potential bias, we generated a corrected sample of eggs by drawing a random number of individuals from each egg mass (using a Poisson distribution with $\lambda=4$; random numbers from this distribution ranged from 0 to 10). We generated five of these corrected samples with replacement ( $n$ ranged from 156 to 183), estimated $N_{\mathrm{b}}$ separately for each (see methods below), and then calculated the harmonic mean of $\hat{N}_{\mathrm{b}}$ across the five samples. We report the mean, bias-corrected $\hat{N}_{\mathrm{b}}$ values, though we found that these were very similar to the values obtained using the entire sample of $\sim 10$ eggs per mass.

Microsatellite genotyping and scoring
Total genomic DNA was extracted from each sample using QIAGEN DNeasy kits (QIAGEN Inc.). Each individual was genotyped at seven microsatellite loci (Table 1). PCR amplifications were run in $20 \mu \mathrm{l}$ volumes with the following components: 100-200 ng genomic DNA, 25 mM $\mathrm{KCl}, 1 \mathrm{mM}$ Tris-HCl pH 9, $0.1 \%$ Triton X-100, 1.5 mM $\mathrm{MgCl}_{2}, 0.2 \mathrm{mM}$ dNTPs, $0.5 \mu \mathrm{M}$ both forward (fluores-cently-labeled) and reverse primers, 0.01 units $/ \mu \mathrm{l}$ of Taq, and water to a final volume of $20 \mu$ l. PCR amplifications were carried out in an MJ Research PTC-200 thermal cycler under the following conditions: $94^{\circ} \mathrm{C}$ for 3 min , followed by 30 cycles of $94^{\circ} \mathrm{C}$ for 30 s , locus-specific annealing temperature (Table 1) for $30 \mathrm{~s}, 72^{\circ} \mathrm{C}$ for 30 s , and a final extension of $72^{\circ} \mathrm{C}$ for 7 min . Microsatellite PCR products were run on an ABI 3730 automated sequencer, and allele sizes were scored using the program GENOTYPER v. 3.7 (Applied Biosystems). The program GENEPOP (Raymond and Rousset 1995) was used to estimate allele frequencies and to test loci for deviations from Hardy-Weinberg equilibrium. We tested all pairs of loci for linkage disequilibrium using the program FSTAT 2.9.3 (Goudet 2002).

Table 1 Microsatellite primer information

| Locus | Primer | Annealing temp ( ${ }^{\circ} \mathrm{C}$ ) | No. of alleles |
| :---: | :---: | :---: | :---: |
| $R P 3^{\mathrm{a}}$ | F: $5^{\prime}$-gaaagcaaaactgggaaagtacata-3' | 50 | 3 |
|  | R: $5^{\prime}$-cctgagagccatccaataagtgeca-3' |  |  |
| RP22 | F: $5^{\prime}$-accccaccagcagaatacaatga- $3^{\prime}$ | 50 | 3 |
|  | R: $5^{\prime}$-agaccagagccagagcaacc- $3^{\prime}$ |  |  |
| RP23 ${ }^{\text {a }}$ | F: $5^{\prime}$-acatagatacaatagatagatagac- $3^{\prime}$ | 45 | 3 |
|  | R: $5^{\prime}$-cacaggaatgtaaaatctggctttc- $3^{\prime}$ |  |  |
| RP193 ${ }^{\text {b }}$ | F: $5^{\prime}$-ccattttctctctgatgtgtgt-3 ${ }^{\prime}$ | 50 | 2 |
|  | R: $5^{\prime}$-tgaagcagatcactggcaaagc-3' |  |  |
| RP385 | F: $5^{\prime}$-attgaaacttgcggetctet-3' | 50 | 2 |
|  | R: $5^{\prime}$-ggcatgtgtccacaatgtaa $-3^{\prime}$ |  |  |
| RP415 | F: $5^{\prime}$-aagtttcattaaagcagatt-3' | 45 | 2 |
|  | R: $5^{\prime}$-ggtatatcttagggttacct-3' |  |  |
| SFC134 ${ }^{\text {b, a }}$ | F: $5^{\prime}$-tgggaaaagactctgtggt-3' | 55 | 3 |
|  | R: $5^{\prime}$-aggaaatgtgtggaagcat- $3^{\prime}$ |  |  |

## Estimates of census population size

We had extensive mark-recapture data for the 2006 season, which allowed us to obtain estimates of the adult population size $(\hat{N})$ using Begon's weighted mean method (Begon 1979) and the program CAPTURE (White et al. 1978). Begon's weighted mean is a modification of the simple Lincoln-Peterson estimate that utilizes capture data from $>2$ trapping occasions. CAPTURE uses maximum likelihood and a model selection procedure to identify the model that best fits the mark-recapture data from among eight possible models. The eight models differ in what variables they include: effects of time on capture, behavioral effects (e.g., "trap-happy" or "trap-shy" behaviors) on capture, and individual variation in capture probability (White et al. 1978). We averaged the estimated population size from the best-fitting model identified by CAPTURE with the estimate obtained using Begon's weighted mean.

Estimates of $N_{\mathrm{b}}$ and $N_{\mathrm{e}}$
There are several methods of estimating $N_{\mathrm{e}}$ indirectly using genetic data, and the time frame to which an estimate applies depends on the method used as well as the sampling design (Waples 2005). In any case, $N_{\mathrm{e}}$ applies to one or more generations, whereas $N_{\mathrm{b}}$ is the effective number of breeding adults in a single reproductive season that produce a single cohort of offspring. $N_{\mathrm{e}}$ is difficult to derive from $N_{\mathrm{b}}$ for organisms with overlapping generations because this requires extensive demographic information about the population (Jorde and Ryman 1995; Waples 2005). However, low estimates of $N_{\mathrm{b}}$ are generally expected to reflect low $N_{\mathrm{e}}$ (Waples 2005). $N_{\mathrm{b}}$ can be estimated by the same methods used to estimate the overall effective size of a population.

Although the 13 microsatellite loci we developed for $R$. pretiosa (Blouin, unpublished data) were polymorphic when surveyed across the species' range (Blouin 2002), only seven proved to be polymorphic in the Crosswater population and none had more than three alleles at a locus. This low level of genetic diversity precluded the use of kinship and pedigree methods to accurately match offspring with their parents or siblings, preventing direct, pedigree-based estimation of $N_{\mathrm{b}}$ (e.g., Araki et al. 2007; Blouin 2003). Consequently, we estimated $N_{\mathrm{b}}$ only through indirect genetic methods.

We compared the estimate of the actual number of breeders obtained from an egg mass count $\left(\hat{N}_{\mathrm{ab}}\right)$ to several $\hat{N}_{\mathrm{b}}$ values obtained from genetic data. The first $\hat{N}_{\mathrm{b}}$ is from a version of the temporal method that uses the differences in allele frequencies between a sample from the adult population and a sample from their offspring (Scribner et al. 1997). We calculated two $\hat{N}_{\mathrm{b}}$ values: one based on allele frequency differences between the adult and egg samples, and another based on allele frequency differences between the adult and metamorph samples. As noted above, each adult-egg estimate of $N_{\mathrm{b}}$ that we report represents the harmonic mean of five random samples generated from the total egg dataset. See Table 2 for notation used.

The first temporal method we used was Waples' (1989) moment-based approach (TM). The standardized variance of allele frequency change for each locus was calculated using Eq. (9) from Waples (1989):
$\hat{F}_{\mathrm{c}}=\frac{1}{K-1} \sum_{i=1}^{K} \frac{\left(x_{i}-y_{i}\right)^{2}}{\left(x_{i}+y_{i}\right) / 2}$
where $K$ is the total number alleles at the locus, $x_{i}$ is the frequency of allele $i$ in the first sample, and $y_{i}$ is the frequency in the second sample. The mean $\hat{F}_{\text {c }}$ across all seven loci was calculated as:

Table 2 Notation used

| $N$ | Actual number of individuals in the population; population census size |
| :--- | :--- |
| $N_{\mathrm{ab}}$ | Actual number of breeding adults |
| $N_{\mathrm{e}}$ | Effective population size |
| $N_{\mathrm{b}}$ | Effective number of breeding adults in one reproductive season |
| $\hat{N}^{2}, \hat{N}_{\mathrm{ab}}, \hat{N}_{\mathrm{e}}, \hat{N}_{\mathrm{b}}$ | Estimates of $N, N_{\mathrm{ab}}, N_{\mathrm{e}}$, and $N_{\mathrm{b}}$ |
| TM | Waples' $(1989)$ temporal moment method of estimating $N_{\mathrm{e}}$ or $N_{\mathrm{b}}$ |
| TL | Temporal likelihood method of $N_{\mathrm{e}}\left(N_{\mathrm{b}}\right)$ estimation from Berthier et al. (2002) |
| LD | Linkage disequilibrium method of $N_{\mathrm{e}}\left(N_{\mathrm{b}}\right)$ estimation |
| $\hat{N}_{\mathrm{b}}^{\mathrm{TM}-\mathrm{Egg}}$ | Temporal moment method estimate of $N_{\mathrm{b}}$, using the adult and egg samples |
| $\hat{N}_{\mathrm{b}}^{\mathrm{TM}-\mathrm{Meta}}$ | Temporal moment method estimate of $N_{\mathrm{b}}$, using the adult and metamorph samples |
| $\hat{N}_{\mathrm{b}}^{\mathrm{TL}-\mathrm{Egg}}$ | Temporal likelihood method estimate of $N_{\mathrm{b}}$, using the adult and egg samples |
| $\hat{N}_{\mathrm{b}}^{\mathrm{TL}-\mathrm{Meta}}$ | Temporal likelihood method estimate of $N_{\mathrm{b}}$, using the adult and metamorph samples |
| $\hat{N}_{\mathrm{b}}^{\mathrm{LD}-\text { Egg }}$ | LD method estimate of $N_{\mathrm{b}}$ from LDNE program, using the egg sample |
| $\hat{N}_{\mathrm{b}}^{\mathrm{LD}-\mathrm{Meta}}$ | LD method estimate of $N_{\mathrm{b}}$ from LDNE program, using the metamorph sample |

$$
\begin{equation*}
\text { mean } \hat{F}_{\mathrm{c}}=\sum\left(K_{j}-1\right) F_{\mathrm{c}_{j}} / \sum\left(K_{j}-1\right) \tag{2}
\end{equation*}
$$

where $K_{j}$ is the number of alleles at locus $j$ and $F_{\mathrm{c}_{j}}$ is the estimate of $F_{\mathrm{c}}$ for locus $j$. Confidence intervals for mean $\hat{F}_{\mathrm{c}}$ were calculated using Eq. (16) from Waples (1989). Because our first sample was collected non-destructively from adults, Waples' (1989) Plan I was the appropriate sampling design. The estimated effective number of breeders was therefore calculated using Eq. (12) from Waples (1989):
$\hat{N}_{\mathrm{b}}=\frac{t}{2\left[\operatorname{mean} \hat{F}_{\mathrm{c}}-\frac{1}{2 S_{\mathrm{o}}}-\frac{1}{2 S_{\mathrm{t}}}+\frac{1}{\hat{N}}\right]}$
where $S_{\mathrm{o}}$ and $S_{\mathrm{t}}$ are sample sizes for the first and second samples, respectively, $t$ is number of generations between the two samples ( 1 in this case), and $\hat{N}$ is the census estimate of the total size of the population from which the $S_{\text {o }}$ sample was drawn (see above for how we obtained $\hat{N}$ ). We designated the adult-egg and adult-metamorph estimates of $N_{\mathrm{b}}$ from this method as $\hat{N}_{\mathrm{b}}^{\mathrm{TM}-\mathrm{Egg}}$ and $\hat{N}_{\mathrm{b}}^{\mathrm{TM}-\text { Meta }}$, respectively.

The second temporal approach we used to estimate $N_{\mathrm{b}}$ was the likelihood-based estimator (TL) of Berthier et al. (2002), implemented in the program TM3. The TL method involves the calculation of likelihoods from coalescentbased gene genealogies and Markov chain Monte Carlo sampling to generate a posterior probability distribution of $N_{\mathrm{e}}$, or in our case, $N_{\mathrm{b}}$. We designated the adult-egg and adult-metamorph estimates of $N_{\mathrm{b}}$ from this method as $\hat{N}_{\mathrm{b}}^{\mathrm{TL}-\mathrm{Egg}}$ and $\hat{N}_{\mathrm{b}}^{\mathrm{TL}-\mathrm{Meta}}$, respectively. A maximum possible $N_{\mathrm{b}}$ value is specified as a Bayesian prior in TM3. Although we did not expect maximum $\hat{N}_{\mathrm{b}}$ to be greater than about 90 frogs (based on the egg mass count), we ran several independent TM3 runs using priors of 200, 300, 400, and 1000 for maximum $N_{\mathrm{b}}$. We set our lowest prior conservatively at

200 to account for the possibility that extra-pair fertilization (i.e., multiple fathers per egg mass) could result in $N_{\mathrm{b}}$ greater than $\hat{N}_{\text {ab }}$. Performing analyses with different priors allowed us to evaluate the sensitivity of $\hat{N}_{\mathrm{b}}^{\mathrm{TL}-\mathrm{Egg}}$ and $\hat{N}_{\mathrm{b}}^{\mathrm{TL}-\text { Meta }}$ to choice of prior. All TM3 analyses were run with 50,000 iterations.

In addition to the two temporal methods, we used the linkage disequilibrium (LD) method to estimate $N_{\mathrm{b}}$ from single samples of eggs and of metamorphs. We designated the $N_{\mathrm{b}}$ estimates from the LD method as $\hat{N}_{\mathrm{b}}^{\mathrm{LD}-\mathrm{Egg}}$ for the egg samples and as $\hat{N}_{\mathrm{b}}^{\mathrm{LD}-\text { Meta }}$ for the metamorph sample. Calculations were performed using the program LDNe (Waples and Do 2007). LDNe incorporates a correction for the bias that is introduced when sample size is less than the true effective size and reports confidence intervals obtained via a new jackknife method (Waples 2006). The mating model for this system is equivalent to monogamy and was selected in the LDNe analyses. We report jackknife confidence intervals for $\hat{N}_{\mathrm{b}}$, with the lowest allele frequency set at 0.05 . By excluding alleles with frequencies $<0.05$, we achieve the most accurate $N_{\mathrm{b}}$ estimate, with an expected tradeoff in precision (Waples and Do 2007). However, even when we ran our analyses with the lowest allele frequency set at 0.01 , confidence intervals were very similar to those obtained when the lowest frequency was set at 0.05 .

Finally, we used the LD method to estimate $N_{\mathrm{e}}$ (as opposed to $N_{\mathrm{b}}$ ) in the adult sample ( $n=176$ ), under a random mating model in LDNe. We acknowledge that there is some uncertainly about how to interpret LD estimates from mixed-cohort samples from species that have overlapping generations (Waples 1991). However, the LD method has become standard for estimating $N_{\mathrm{e}}$ from such samples (e.g., Aspi et al. 2008; Durrant et al. 2008), so our data should still be useful for comparative purposes.

## Results

## Genetic diversity

Expected heterozygosity $\left(H_{\mathrm{e}}\right)$ for the seven microsatellite loci in this R. pretiosa population was 0.40 as calculated from the adult sample. The maximum number of alleles per locus was three. Only one locus in one sample (RP3 in the metamorph sample) was barely out of Hardy-Weinberg equilibrium ( $P=0.0071$; Bonferroni-corrected nominal value of 0.00714 ). We found one locus pair (RP3 $\times$ RP385) with barely significant linkage disequilibrium in the adult sample ( $P=0.00238$; Bonferroni-corrected nominal value of 0.002381 ). About six pairs of loci in the metamorph sample and three to four pairs in each of the five random egg samples exhibited significant linkage disequilibrium (data not shown).

Estimates of census population size
The two methods of population size estimation yielded very similar results. Begon's weighted mean method gave $\hat{N}$ of 444 (95\% C.I. 343-545). CAPTURE identified the $\mathrm{M}_{\mathrm{t}}$ model as the most appropriate for our mark-recapture data. Under this model, each individual has the same probability of capture on a given trapping occasion, but these probabilities are variable across trapping occasions (White et al. 1978). $\hat{N}$ from the CAPTURE analysis was 412 ( $95 \%$ C.I. 343-513). The average of the two $\hat{N}$ values is 428 ( $95 \%$ C.I. 343-529).

Estimates of $N_{\mathrm{b}}$ and $N_{\mathrm{e}}$
Point estimates of $N_{\mathrm{b}}$ from the TL analysis using the program TM3 were insensitive to the value of Bayesian prior for maximum $N_{\mathrm{b}}$ (Table 4). As one might expect, the upper confidence limit did increase with increasing prior. However, even if extra-pair fertilization was rampant in this population, such that the number of breeding males was more than twice the number of breeding females, maximum $N_{\mathrm{b}}$ should not exceed 200 . Thus, using 200 as the upper prior for our reported values (Table 3) probably produced overly liberal upper confidence intervals, even if we are confident in the point estimates.

Estimates of $N_{\mathrm{b}}$ are presented in Table 3. Doubling the number of egg masses found in the 2006 breeding season resulted in an estimate of 90 breeding adults $\left(\hat{N}_{\mathrm{ab}}=90\right)$. The temporal methods (TM and TL) yielded similar point $\hat{N}^{\text {estimates }}$ of $N_{b}$ for the adult-egg $\left(\hat{N}_{\mathrm{b}}^{\mathrm{TM}-\mathrm{Egg}}=65.0\right.$, $\left.\hat{N}_{\mathrm{b}}^{\mathrm{TL}-\mathrm{Egg}}=87.3\right)$ and adult-metamorph comparisons $\left(\hat{N}_{\mathrm{b}}^{\mathrm{TM}-\text { Meta }}=82.5, \hat{N}_{\mathrm{b}}^{\mathrm{TL}-\text { Meta }}=117.2\right)$. Estimates of $N_{b}$ from the LD method were also very similar between the two life stages $\left(\hat{N}_{b}^{L D-E g g}=68.5, \hat{N}_{b}^{L D-M e t a}=56.2\right)$. Thus,

Table 3 Estimates of effective number of breeders $\left(N_{\mathrm{b}}\right)$ and $N_{\mathrm{b}} / N$ in the CW population

| Method |  | Estimate | 95\% C.I. | $N_{\mathrm{b}} / N$ |
| :---: | :---: | :---: | :---: | :---: |
| Temporal methods |  |  |  |  |
| Adult-egg |  |  |  |  |
| TM | $\hat{N}_{\mathrm{b}}^{\mathrm{TM}-\mathrm{Egg}}$ | 65.0 | 18-195 | 0.15 |
| TL | $\hat{N}_{\mathrm{b}}^{\mathrm{TL}-\mathrm{Egg}}$ | 86.4 | 20-200 | 0.20 |
| Adult-metamorph |  |  |  |  |
| TM | $\hat{N}_{\text {b }}^{\text {TM-Meta }}$ | 82.5 | 23-252 | 0.19 |
| TL | $\hat{N}_{\mathrm{b}}^{\text {TL-Meta }}$ | 117.2 | 27-200 | 0.27 |
| Linkage disequilibrium methods |  |  |  |  |
| Egg |  |  |  |  |
| LD | $\hat{N}_{\mathrm{b}}^{\mathrm{LD}-\mathrm{Egg}}$ | 68.5 | 30-108 | 0.16 |
| Metamorph |  |  |  |  |
| LD | $\hat{N}_{\text {b }}^{\text {LD-Meta }}$ | 56.2 | 26-108 | 0.13 |

Estimates are given for two temporal methods: Waples' (1989) temporal moment (TM) and the temporal likelihood method (TL) of Berthier et al. (2002). Estimates from the linkage disequilibrium method were obtained using LDNE (Waples and Do 2007). Estimates from these various methods are listed along with their $95 \%$ confidence intervals (Bayesian credible intervals for the TL estimates) and $N_{\mathrm{b}} / N$ ratios

Table 4 Harmonic means of $\hat{N}_{\mathrm{b}}$ values for four choices of Bayesian prior for maximum $N_{\mathrm{b}}$

| Prior | Harmonic mean | Lower C.L. | Upper C.L. |
| :--- | :--- | :--- | :--- |
| 200 | 86.35 | 20 | 200 |
| 300 | 83.37 | 17 | 280 |
| 400 | 91.04 | 19 | 352 |
| 1000 | 88.61 | 15 | 519 |

Means were calculated from the five bias-corrected egg samples for each prior
we see (1) point estimates from the egg stage ( $65.0,87.3$, and 68.5) that are fairly close to the simple estimate of 90 breeding adults, and (2) no indication of a massive drop in $N_{\mathrm{b}}$ in going from the egg to metamorph stage.

The LD estimate of $N_{\mathrm{e}}$ in the adult sample was 36.7 (95\% C.I. 19-71.9). Thus, the best point estimate of $N_{\mathrm{e}} / N$ for this population $=36.7 / 428=0.086$.

## Discussion

Esitmates of $N_{\mathrm{b}} / N$ across the various methods ranged from 0.13 to 0.27 (Table 3). These values are similar to those found for Rana temporaria populations in Finland (0.060.17; Schmeller and Merila 2007) and Britain (0.3330.365 ; Brede and Beebee 2006), but considerably higher than those of toad (Bufo bufo) populations in Britain (0.007-0.012 Scribner et al. 1997; 0.034-0.040 Brede and

Beebee 2006). The $N_{\mathrm{e}} / N$ ratio estimated for the adult population was 0.086 , which again is in the general range of DNA-based estimates for other ranid frogs (Hoffman et al. 2004; Schmeller and Merila 2007). Thus, our data are consistent with previous suggestions that $N_{\mathrm{e}} / N$ ratios in ranid frogs are in the typical range for vertebrates (e.g., $\sim 0.1-0.4$ ), while those for bufonids are much lower (Hoffman et al. 2004; Brede and Beebee 2006).

Our main objective was to test a hypothesis about what features of the life cycle of $R$. pretiosa cause $N_{\mathrm{e}}$ to be reduced relative to $N$. By obtaining separate $N_{\mathrm{b}}$ estimates using egg and metamorph samples we could determine if these $N_{\mathrm{b}}$ estimates differed from each other and from the simple estimate from counting egg masses $\left(\hat{N}_{\mathrm{ab}}\right)$. To our knowledge, this study is the first to take such an approach. We found that: (1) Estimates for $N_{\mathrm{b}}$ at the egg stage using both temporal methods and the LD method did not differ dramatically from $\hat{N}_{\mathrm{ab}}=90$; and (2) estimates for $N_{\mathrm{b}}$ were similar for eggs and metamorphs (Table 3). The first result is consistent with what we would expect to find if each female produced a single egg mass, each egg mass was fertilized by a single male, and each male bred with only one female. The second result suggests little non-random (family-based) mortality occurred between egg laying and metamorphosis.

The first result was expected because, like many ranid frogs, $R$. pretiosa females are thought to lay one egg mass per season (Olson and Leonard 1997) and explosive breeding reduces the opportunity for males to mate with multiple females (Wells 1977). Indeed, in this year all breeding occurred on a single night. On the other hand, sex ratios in breeding populations are male-biased, which could promote extra-pair fertilization, where some egg masses are fertilized by more than one male. This could occur either passively by free-swimming spermatozoa in communal breeding areas (Laurila and Seppa 1998) or actively by lone 'pirate' (or 'sneaker') males that fertilize some of the eggs of breeding pairs (Vieites et al. 2004). If extra-pair fertilization was frequent, the resulting decrease in the variance of male reproductive success would increase $N_{\mathrm{b}}$ estimates at the egg stage relative to $\hat{N}_{\mathrm{ab}}$ (Sugg and Chesser 1994). In our observations of hundreds of breeding pairs of spotted frogs over multiple years in this population, we have witnessed few instances of behavior that would suggest the occurrence of 'clutch piracy'. Thus, we interpret our results as consistent with predictions of a basically monogamous mating system in which each female lays a single clutch per year. One practical consequence of these results is that they support the use of egg mass counts as a cost-effective method of population monitoring, in that they probably do give a reasonable estimate of the number of adults that bred in a given year. Whether egg mass counts can consistently provide reliable estimates of $N_{\mathrm{b}}$
depends on how typical are the results that variance in family size apparently increased little after egg laying.

If our results for this population in 2006 are typical for $R$. pretiosa, then the reduction of $N_{\mathrm{e}}$ relative to $N$ in this species is not owing to the inflation of variance in family size that occurs between the egg and metamorph stages. Thus, we might consider other factors such as year-to-year fluctuations in population size. Of course, our results are from one year in a single population and may not be typical. Water levels in the pond were very high in 2006, which may have contributed to unusually high survival of entire egg masses. Such an environmental effect was also noted by Schmeller and Merila (2007), who suggested that high egg-to-metamorph mortality during a short growing season may have been responsible for low $N_{\mathrm{b}} / N$ ratios in two populations of $R$. temporaria. Together, these observations suggest the interesting hypothesis that the $N_{\mathrm{b}} / N$ ratio varies from year to year (or from population to population) depending on habitat quality. Indeed, $N_{\mathrm{b}} / N$ and $N_{\mathrm{e}} / N$ might even be predictable from environmental measurements.

This study is the first attempt to determine where in the ranid life cycle $N_{\mathrm{b}}$ (and by extension, $N_{\mathrm{e}}$ ) is reduced. More studies will be needed before a consensus is reached about the importance of different factors. Here we provide some of the first data on the subject, and suggest the hypothesis that $N_{\mathrm{b}} / N$ might vary substantially in time and space owing to habitat conditions that influence the survival of eggs and larvae. The approach of estimating $N_{\mathrm{b}}$ using genetic data from a single cohort at more than one life stage should prove valuable in future studies on the determinants of effective population size in amphibians and other taxa.

Acknowledgments This research was conducted as part of Ivan Phillipsen's PhD dissertation. We would like to thank Patty Stenberg for help in tissue collection. Robin Waples provided suggestions on the methodology and helpful comments on an early draft of the manuscript. Two anonymous reviewers also provided useful input on the manuscript. We thank the Oregon State University Center for Genome Research and Biocomputing (CGRB) for assistance with microsatellite genotyping.

## References

Araki H, Waples RS, Ardren WR, Cooper B, Blouin MS (2007) Effective population size of steelhead trout: influence of variance in reproductive success, hatchery programs, and genetic compensation between life-history forms. Mol Ecol 16:953-966
Aspi J, Roininen E, Kiiskilä J, Ruokonen M, Kojola I, Bljudnik L, Danilov P, Heikkinen S, Pulliainen E (2008) Genetic structure of the northwestern Russian wolf populations and gene flow between Russia and Finland. Conserv Genet. doi:10.1007/s10592-008-9642-x. http://dx.doi.org/10.1007/s10592-008-9642-x
Begon M (1979) Investigating animal abundance: capture-recapture for biologists. University Park Press, Baltimore

Berthier P, Beaumont MA, Cornuet J-M, Luikart G (2002) Likeli-hood-based estimation of the effective population size using temporal changes in allele frequencies: a genealogical approach. Genetics 160:741-751
Berven KA (1995) Population regulation in the wood frog, Rana sylvatica, from three diverse geographic localities. Aust J Ecol 20:385-392
Berven KA, Grudzien TA (1990) Dispersal in the wood frog (Rana sylvatica): implications for genetic population structure. Evolution 44(8):2047-2056
Blouin MS (2002) Genetic data for recovery of the oregon spotted frog, Rana pretiosa, in the Western United States, final report for USGS FY 2000 species at Risk Program project. USGS, Corvallis, Oregon
Blouin MS (2003) DNA-based methods for pedigree reconstruction and kinship analysis in natural populations. Trends Ecol Evol 18:503-511
Brede EG, Beebee TJ (2006) Large variations in the ratio of effective breeding and census population sizes between two species of pond-breeding anurans. Biol J Linn Soc 89:365-372
Briggs JL (1987) Breeding biology of the cascade frog, Rana cascadae, with Comparisons to $R$. aurora and $R$. pretiosa. Copeia 1987:241-245
Crouch WB, Paton PWC (2000) Using egg-mass counts to monitor wood frog populations. Wildl Soc Bull 28:895-901
Crow JF, Morton NE (1955) Measurement of gene frequency drift in small populations. Evolution 9:202-214
Cushman KA, Pearl CA (2007) A conservation assessment for the Oregon spotted frog (Rana pretiosa). USDA Forest Service and USDI Bureau of Land Management, Oregon
Durrant C, Beebee T, Greenaway F, Hill D (2008) Evidence of recent population bottlenecks and inbreeding in British populations of Bechstein's bat, Myotis bechsteinii. Conserv Genet. doi:10.1007/ s10592-008-9639-5. http://dx.doi.org/10.1007/s10592-008-9639-5
Easteal S (1985) The ecological genetics of introduced populations of the giant toad Bufo marinus. II. Effective population size. Genetics 110:107-122
Frankham R (1995) Effective population size/adult population size ratios in wildlife: a review. Genet Res 66:95-107
Funk WC, Blouin MS, Corn PS et al (2005) Population structure of Columbia spotted frogs (Rana luteiventris) is strongly affected by the landscape. Mol Ecol 14:483-496
Goudet J (2002) FSTAT version 2.9. 3.2, updated from Goudet (1995) FSTAT: a computer program to calculate $F$-statistics. J Hered 86:485-486
Hayes MP (1997) Status of the Oregon spotted frog (Rana pretiosa) in the Deschutes Basin and selected other systems in Oregon and northeastern California with a rangewide synopsis of the species’ status. Nature Conservancy under contract to the US Fish and Wildlife Service with assistance from the Oregon Department of Fish and Wildlife, Portland final report
Hill WG (1981) Estimation of effective population size from data on linkage disequilibrium. Genet Res 38:209-216
Hoffman EA, Schueler FW, Blouin MS (2004) Effective population sizes and temporal stability of genetic structure in Rana pipiens, the northern leopard frog. Evolution 58:2536-2545
Jorde PE, Ryman N (1995) Temporal allele frequency change and estimation of effective size in populations with overlapping generations. Genetics 139:1077-1090
Laurila A, Seppa P (1998) Multiple paternity in the common frog (Rana temporaria): genetic evidence from tadpole kin groups. Biol J Linn Soc 63:221-232
Leonard W, Hallock L, McAllister KR (1997) Natural history notes: Rana pretiosa (Oregon spotted frog). Behavior and reproduction. Herpetol Rev 28:86

Licht L (1969) Comparative breeding behavior of the red-legged frog (Rana aurora aurora) and the western spotted frog (Rana pretiosa pretiosa) in southwestern British Columbia. Can J Zool 47:1287-1299
Licht LE (1971) Breeding habits and embryonic thermal requirements of the frogs, Rana aurora aurora and Rana pretiosa pretiosa, in the Pacific Northwest. Ecology 52:116-124
McAllister KR, Leonard WP (1997) Washington state status report for the Oregon spotted frog, p. 38. Washington Department of Fish and Wildlife, Olympia
Merrell DJ (1968) A comparison of the estimated size and the "effective size" of breeding populations of the leopard frog, Rana pipiens. Evolution 22:274-283
Monsen KJ, Blouin MS (2003) Genetic structure in a montane ranid frog: restricted gene flow and nuclear-mitochondrial discordance. Mol Ecol 12:3275-3286
Nussbaum RA, Brodie ED Jr, Storm RM (1983) Amphibians and reptiles of the Pacific Northwest. University of Idaho Press, Moscow
Olson DH, Leonard WP (1997) Amphibian inventory and monitoring: a standardized approach for the PNW. In: Olson DH, Leonard WP, Bury RB (eds) Sampling amphibians in lentic habitats. Society for Northwestern Vertebrate Biology, Olympia, pp 1-21
Raymond M, Rousset F (1995) GENEPOP (version 1.2): population genetics Software for exact tests and ecumenicism. J Hered 86:248-249
Rowe G, Beebee T (2004) Reconciling genetic and demographic estimators of effective population size in the anuran amphibian Bufo calamita. Conserv Genet 5:287-298
Schmeller DS, Merila J (2007) Demographic and genetic estimates of effective population and breeding size in the amphibian Rana temporaria. Conserv Biol 21:142-151
Scribner KT, Arntzen JW, Burke T (1997) Effective number of breeding adults in Bufo bufo estimated from age-specific variation at minisatellite loci. Mol Ecol 6:701-712
Seburn D, Seburn C (2000) Conservation priorities for the amphibians and reptiles of Canada. Prepared for World Wildlife Fund Canada and the Canadian amphibian and reptile conservation network
Stuart SN, Chanson JS, Cox NA et al (2004) Status and trends of amphibian declines and extinctions worldwide. Science 306: 1783-1786
Sugg DW, Chesser RK (1994) Effective population size with multiple paternity. Genetics 137:1147-1155
US Fish and Wildlife Service (2005) Species assessment and listing priority assessment form. Western Washington Fish and Wildlife Office, Lacey. Available at http://ecos.fws.gov/docs/candforms_ pdf/r1/D02A_V01.pdf
Vieites D, Nieto-Román S, Barluenga M et al (2004) Post-mating clutch piracy in an amphibian. Nature Rev Genet 4:598-612
Waples RS (1989) A generalized approach for estimating effective population size from temporal changes in allele frequency. Genetics 121:379-391
Waples RS (1991) Genetic methods for estimating the effective size of cetacean populations. Rep Intern Whal Comm 13:279-300 (special issue)
Waples RS (2005) Genetic estimates of contemporary effective population size: to what time periods do the estimates apply? Mol Ecol 14:3335-3352
Waples RS (2006) A bias correction for estimates of effective population size based on linkage disequilibrium at unlinked gene loci. Conserv Genet 7:167-184
Waples RS, Do C (2007) LDNE: a program for estimating effective population size from data on linkage disequilibrium. Mol Ecol Resour 8:753-756

Watson JW, McAllister KR, Pierce DJ, Alvarado A (2000) Ecology of a remnant population of Oregon spotted frogs (Rana pretiosa) in thurston county, Washington. Washington Department of Fish and Game, Olympia

Wells K (1977) The social behaviour of anuran amphibians. Anim Behav 25:666-693
White GC, Burnham KP, Otis DL, Anderson (1978) Users manual for program. CAPTURE Utah State University Press, Logan


[^0]:    I. C. Phillipsen $(\boxtimes)$

    Department of Zoology, Oregon State University, 3029 Cordley Hall, Corvallis, OR 97331-2914, USA
    e-mail: philliiv@science.oregonstate.edu
    J. Bowerman

    Sunriver Nature Center \& Observatory, P.O. Box 3533, Sunriver, OR 97707, USA
    M. Blouin

    Department of Zoology, Oregon State University, 3029 Cordley Hall, Corvallis, OR 97331-2914, USA

